

Water and Soil Monitoring for the Protection of Environment and Human Health

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ABSTRACT

The first pillar of the protection of the environment and also, as a positive consequence, of the human community living in this environment, is the establishment of protective monitoring programs. Current monitoring programs for water and soil are based on sampling and laboratory analysis of chemical and microbiological variables. Parallel to this traditional approach, methods to measure effects directly on living organisms, at both individual and community level, have been integrated into monitoring plans. The present paper reviews the state of the art of the research activities in the field of water and soil monitoring carried out by the Italian National Research Council (CNR) Institutes: this review is the outcome of a survey conducted by the Working Group 2 established in the framework of the CNR Environment and Health Inter-departmental Project, PIAS CNR. Emerging problems, such as the presence of nanoparticles and perfluorinated compounds in the environment, and future developments of the monitoring techniques are also discussed.

1. INTRODUCTION

The protection of the environment and human health from toxic agents is traditionally based on the selection of a list of potentially dangerous substances or agents and the statement of the corresponding emission limit values or quality standards.

Methodologies to establish quality standards are based on physicochemical and toxicological data which are normally collected and organized in a risk assessment document. The prioritization procedure used to establish the list of pollutants is based on the knowledge of the toxicological properties and data on use and production amounts. In order to be validated, all these procedures need a large amount of data which are actually not available for the millions of synthetic molecules prepared and brought into the market.

As a consequence, the approach, based on

emission limit values at the discharge and on environmental quality standards in the receptor compartment, could not be really protective for the environment.

A step forward in the legislation on water quality protection was the publication of the Water Framework Directive (WFD, Directive 2000/60/EC) that introduced the concept of water bodies protection: it moved from a regulation based on emission control to one that is based on the protection of the ecological quality of receiving water bodies. On the assumption that the repression of discharge by imposing emission limit is not sufficient to protect the environment, it is crucial to verify whether the receiving body is able to support activities which are imposed on it, keeping as much intact as possible the ecological community that resides there. The aim of the WFD is to achieve a good quality status from the ecological point of view, namely to ensure that all bodies

could support a biodiverse ecological community.

In parallel with this innovative approach based on ecological classification, in order to control the chemical pollution, WFD establishes a priority list of compounds which, for production volumes and/or use and hazardous characteristics in terms of toxicity, persistence and bioaccumulation, pose a risk to the aquatic ecosystem or to human health. The pollution control is based on a combined approach which sets limits on emissions (left to Member States) and maximum allowable levels in the receiving water body, expressed as environmental quality standards (EOS) which are fixed in a recently adopted EC Directive (105/2008/EC).

The use of living organisms and their community as monitoring tools has many advantages. Organisms, living in the environment under study, are constantly exposed to the physical, biological and chemical influences of that environment. Organisms can often accumulate significant quantities of compounds even if exposed to very low concentrations in the environment.

It is nevertheless difficult to correlate the measured adverse effects on the ecological community with the presence of specific classes of chemical compounds because also the hydromorphological and physicochemical alteration of the natural habitat influences the structure of the community. In order to detect sub-lethal effects, single living organisms are the best indicators of environmental alteration: if chemical tests only detect "known" substances, the measuring of effects on organisms by appropriate biomarkers can highlight not only the biological response to unknown substances but can also evidence the synergistic effects that may be caused by a mix of different substances.

Integrated monitoring systems are the most effective tool to highlight the interactions among substances by pointing out the responses at different levels: responses which can be so lethal that they affect the composition of the community, or sub -lethal responses that act on the bodymetabolic-physiological models or interact with genetic transmission mechanisms.

Low-cost or highly innovative technologies have been developed for the application in field/in situ (spectroscopic or sensing). in order to achieve a rapid characterization of the site with a high spatial/temporal resolution: in alternative to chemical or physicochemical monitoring systems. techniques based on the measurement of biological response have been proposed, as they can evaluate the possible risk, even in the absence of a direct chemical measurement of a particular pollutant. The role of biomarkers, possibly integrated in a biosensor system, is crucial for the development of an early warning system which could prevent adverse effects on the ecological community and human health.

2. Working methodology in working group 2 of the CNR Environment and Health Inter-Departmental Project (PIAS)

The working group 2 (WG2) on "Monitoring Systems for Soil and Water", established in the framework of the PIAS-CNR Project, has defined its field of interest in the development and use of monitoring techniques, technologies or innovative methods for soil and water monitoring, where a situation of environmental pollution represents a potential risk for human health.

After a preliminary review of literature focused on the various techniques developed for the monitoring of soils and waters, WG2 researchers were invited to put together their scientific expertise and produce a state of the art report of the activities regarding the monitoring of environmental impacts which could be relevant for human health. Assuming the central role of the integrated approach, a review of the different professional profiles present in the CNR Institutes was carried out, to facilitate the creation of a multidisciplinary group of researchers and the sharing of different expertise ranging from instrumental analytical chemistry to ecological assessment.

3. State of the art of CNR activities

Ecological risk assessment (ERA) has been defined as "the practice of determining the nature and likelihood of effects of anthropogenic actions on animals, plants, and the environment" (1). A correct analysis of the complex interactions between the pollution caused by humans and the environment requires the application of a multidisciplinary approach and the determination of different parameters, that can describe the exposure levels and convert them into individual warning situations (2).

A clear example of this operating method is described by the Triad approach used in the ecological risk assessment of sediments (2,3) and allows the investigation of the possible negative effects of toxic chemicals at different levels of biological organization, from single organism to population and/or community level (2). The Triad paradigm enables the assessment of potentially hazardous effects on ecosystems by considering simultaneously chemical concentration, bioavailability of pollutants and the ecotoxicological profile of the environmental matrix under observation. The latter is usually determined by a set of ecotoxicological tests as well as by monitoring possible ecological alterations, quantified by changes in different structural and functional community attributes. This integrated approach on different levels of monitoring should be adopted to provide full details of the impact on both the water and the soil in the site of interest. Methods that are relevant at different levels of specificity are needed because some parameters (such as biomarkers) describe effects at suborganism levels of biological organization (4) and different phases of stress syndrome evolution in model organisms (5,6). On the contrary, other ecotoxicological endpoints (e.g. survival and reproduction) indicate possible direct effects at population level (7). Chemical analyses reveal the presence of potentially dangerous substances in soils, but cannot be used to quantify bioavailable fractions (8) that play a more relevant role in threatening ecosystem integrity (9-11). Finally, a direct evaluation of community structure and functionality should clearly detect overall environmental effects on ecosystems (12).

As everybody knows, biomarkers have been defined as sublethal responses to environmental chemicals at different levels of biological organization (e.g, molecular, cellular, tissutal, physiological, behavioral, of organisms) which evidence an alteration respect to the natural status (13,14).

Toxic effects caused by exposure to environmental pollutants can alter endpoints at different levels of biological organization (4,15) (Fig. 1).

In particular, the classical biological tools applied in ERA (i.e, bioassays and ecological parameters) are able to highlight the impairments from the organism to the population–community level. This analytical system cannot, however, be used to investigate early effects in organisms exposed to pollutants (i.e, from early

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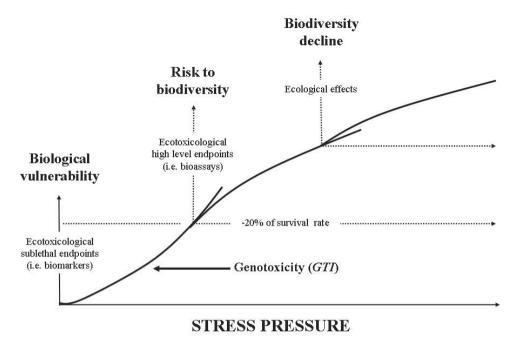


Figure 1: Different biological toxic effects induced by environmental contamination. From Dagnino et al. (15).

sublethal stress syndrome to the onset of reduced survival). The investigation of the initial phases of biological impairment plays a crucial role in determining the vulnerability level reached by the biotic resources in those cases where no evident changes in the traditional, high level endpoints are detected. Therefore, in order to complete the analysis of the spectrum of possible biological effects induced by environmental contamination, the alterations in sublethal endpoints measured on sentinel organisms can be evaluated (6). In spite of the high sensitivity of these types of parameters, it must be stressed that it is possible to clearly infer the stress syndrome degree in organisms exposed to toxic chemicals by the application of ad hoc models. This is done by using a complete battery of biomarkers at different levels of biological organization (i.e., molecule, cell, tissue, organ, organism) on model organisms (5).

Therefore an extensive monitoring requires

different skills, ranging from purely analytical capacities for the identification of compounds present in the matrix under investigation, up to bio-molecular techniques to assess effects of pollutants on the gene expression.

In the following paragraphs, a survey on the activities of CNR institutes will be presented, covering many of the skills described above and needed for the establishment of an effective monitoring project. The covered field of expertise ranges from advanced chemical analysis, in laboratory as well as field research, to traditional ecotoxicological assays, biomarker assessment in exposed and natural organisms, up to studies about the alterations of structure and function of ecological communities.

3.1 Laboratory chemical analysis

The off-site instrumental techniques are characterized by manual sample collection and transfer in centralized laboratory units where advanced analytical equipment is available. Many chemical monitoring activities with instrumental methods are currently on-going in CNR Institutes, also in response to the Water Framework Directive (WFD, 2000/60/EC) requirements.

Several groups are involved in monitoring metals in Italian soil, surface and groundwater. Metals present in soil and waters can accumulate in the trophic chain and represent a risk for the final consumers. The Tuscany region in Central Italy, due to geological and historical reasons (past mining activities present in the territory since the Etruscan era), is particularly impacted by metals and many CNR groups are working on metals contamination in this area.

Trace metal concentrations were monitored in some urban soils of three medium sized towns of coastal Tuscany. (16) Soil samples were collected in roadsides, urban agricultural soils (allotments), playgrounds and public parks. The analysis included total metal content (Pb, Cu, Zn, Ni, Cd), and sequential extraction. Lead reached the highest levels in the soils and was higher near roads. In urban agricultural soils and in allotments Cu was present in noticeable quantities (300 mg/kg). The presence of Cu in urban soils seems to be typical of soils used for a long period as agricultural land, especially vineyards in the area covered by this study. Sequential extractions were performed to evaluate the mobility of the metals and to better understand the impact of the anthropogenic activity on urban sites.

Mercury contamination in the Cecina river basin (Tuscany, Italy) has been studied by Scerbo et al. (17). Mercury was measured in waters, sediments and fish of the river and its most important tributaries. In fish samples the organometallic metabolites of mercury were also determined. Particularly high concentrations were found in the sediments of the S. Marta channel flowing into the Cecina, where a chloro-alkali plant discharges its wastes, and high levels were still detectable 31 km downstream from the confluence, bioaccumulating also in fish species.

Italy, and particularly Tuscany, is strongly interested bv boron contamination because of the presence in its territory of active volcanism, geothermal activity and mineralized areas. The compliance with the EU normative is technically complex and economically very expensive. The limit of 1 mg/l imposed by the European in drinkable waters Union for boron (98/93/EC) is based on the "precautionary principle", considered that the effect of boron on humans is at present poor and contradictory.

New geochemical and isotopic investigations on waters (δD , $\delta^{18}O$, $\delta^{11}B$, $\delta^{13}C$, $\delta^{15}N$, ${}^{87}Sr/{}^{87}Sr$) and soils ($\delta^{11}B$, ${}^{87}Sr/{}^{87}Sr$) were carried out in Southern Tuscany where boron anomalies occur and the assumption of this element through drinking water or agricultural products could have an adverse effect on the health of the local population (18-21).

determination The of abnormal concentrations of trace metals in soil is crucial to highlight the possible presence of contamination. These abnormalities are identified on the basis of the knowledge of natural concentrations, expressed in terms of "background" (natural background levels) or "baseline" (currently found contents). The criteria for the determination of these "natural" concentrations have been the subject of intense international debate for many years. It is therefore necessary to evaluate the potential of other pollution indicators from diffuse sources, complementary to the existing ones (soils and sediments). In order to achieve this, the properly standardized use of higher plants offers a promising tool to establish a precise date for the event and to assess the spatial extension of the contamination (23-25).

In parallel to soil and groundwater studies, presence, distribution and accumulation of metals in sea areas were investigated: samples of Mytilus galloprovincialis were collected monthly during the July 1999-June 2000 period from two mussel culture areas influenced by urban and industrial wastes (26). These stations, subject to different environmental impact conditions, are located in the coastal area of Taranto Gulf (Ionian Sea, Southern Italy). Metals (Cd, Cu, Pb, Zn, Fe and As) were determined by atomic absorption spectrophotometry (AAS) in the whole soft tissue of mussels. Seasonal changes in metal concentrations were observed. Metals exhibited maximum values in later winter-early spring, followed by a progressive decrease during summer.

Metal concentrations were similar to those detected in other Italian coastal zones, and indicate that the seafood under investigation poses no hazard to human health because metal content is within the permissible range established for safe consumption by humans.

For many years the presence of toxic inorganic fibrous particles such as asbestos in drinking waters has been of great concern for their direct impact on human health.

The assessment of the diffusion of inorganic fibrous particles in the environment is performed through detection, identification and quantification of mineral inorganic particles present in animal and human tissues and biological fluids from impacted areas. In these same areas, a comparison is carried out among the types and amounts of particles detected in the biological samples and the fibers present in aerosol and mineral outcrop. An evaluation of the biological effects of some of the fibrous mineral phases present in the rocks is also performed (27-30).

In the field of the advanced methodologies for the monitoring of organic micropollutants in water, the CNR Water Research Institute (CNR-IRSA) plays a well acknowledged role. Besides the determination on "classical" persistent and priority organic pollutants (chlorinated pesticides, PCB, PAH, alkylphenols), advanced analytical methods based on LC-MS technique are under development for the determination of metabolites and emerging pollutants in various surface, drinking and ground waters in Italy (31-33).

Emerging substances are those compounds or groups of compounds produced or used in significant quantities but which are not currently restricted by regulation due to the lack of information about their effective environmental diffusion and toxicity. In this category, many substances are polar substances such as perfluorinated compounds, PPCP (pharmaceutical and personal care products), estrogens. The CNR-IRSA research group is also the national focal point of an international network,NORMAN(Network of Reference Laboratories for Emerging substances) and the one responsible for the substance prioritization procedures at European level under the Water Framework Directive Common Implementation Strategy.

3.2 Biomarker-bioassay

Bioassays are typically used to measure the effects of some substance on a living organism.

They can be classified on the basis of the type of response, by discriminating

the end-point level: a high response level is associated to the survival rate or reproduction inhibition, while a low response level is connected to sublethal effects that are revealed by specific physiological or genomic alterations. Among the former type, the most used ecotoxicological tests are those based on the measurement of EC50 (Effect Concentration) or LC50 (Lethal Concentration), i.e. the concentrations which exert an effect on 50% of the organisms under test. The latter type of bioassays is based on the measurement of specific biomarkers: experimental tests evaluate effects that are not lethal for the organism, like a change in a metabolic protein or a behavioral modification. As indicated by Dagnino et al. (15), the separate evaluation of different levels of response helps to avoid a misinterpretation of the ecotoxicological results.

At the CNR Marine Science Institute (CNR-ISMAR), studies are carried out about the health condition assessment of marine invertebrates and vertebrates in relation to environmental stressors. (34-36) Moreover, there is a great interest about the deployment of biochemical, histo-cytochemical and histopathological biomarkers as early warning systems in coastal marine environments monitoring that is the object of the research at the CNR Institute of Biomedicine and Molecular Immunology (CNR-IBIM) Cell Stress and Environment Research Unit (37-46). This research group has contributed to the identification of cellular and molecular stress markers in the marine environment by using the sea urchin as a model system. This is a common organism in our shores and has a great ecological and commercial importance. Laboratory as well as field experiments showed that, when responding to chemical (heavy/essential metals) or physical (temperature, acidity, ionizing radiations) stressors, the adult immuno-competent cells and embryos or larvae of the sea urchin express specific markers, better known as stress proteins, including heat shock proteins (hsp70) (39-41), metallothionein, (38) and acetylcholinesterase (37). Other studies have shown that environmental stress causes DNA damage in the form of broken singleand double-strand (44), and variations in the levels of other stress and apoptotic markers in response to exposures to heavy/essential metals (cadmium/manganese) and/or UV-B/X radiation (42,43). Results at cellular and molecular (proteins and mRNA) levels from laboratory exposures were compared to those obtained using samples collected in field studies carried out in the Mediterranean and Northern seas in order to bridge together field ecology and laboratory-oriented molecular toxicology (45.46). Most of the markers tested were sensitive to the stress conditions used. The results of the research support the suitability of sea urchin cells and embryos as valid tools to bio-monitor the effects of physical and chemical stress on marine aquatic ecosystems.

At the CNR Biophysics Institute (CNR-IBF) studies on the effects of environmental pollutants in eukaryotic microorganisms are ongoing; specific type of physiological, cellular and molecular level responses have been identified in the presence of environmental pollutants (47-52). This kind of studies is a valuable complement to chemical analysis, as it can provide valuable information on the potential toxicity to the organs, in order to detect the first symptoms of exposure. The effects studied include:

- the change in the intracellular pool of non-protein thiols (glutathione and phytochelatins) in phytoplankton algae to be used as a biomarker of heavy metals bioavailability.

- the variation of photosynthetic microorganisms in aquatic and motility in response to exposure to environmental pollutants.
- short-term genotoxicity tests on cell cultures of specific strains of yeast used as a model system.

An integrated approach of biological assays performed with different microorganisms can be applied to water and sediment elutriate collected in impacted coastal and inland areas, such as estuaries, ports, urban areas.

In vivo measures by micro-spectroscopy and micro-spectrofluorimetry of the photosynthetic compartment of planktonic species, present in water contaminated by organic compounds and/or heavy metals, can determine the effects of these contaminants on the composition of pigments. the ratio chlorophyll: photosynthetic carotenoids. efficiency (53-56). Toxic algal species identification and quantification could be done using techniques of optical microscopy, and/or fluorescence for taxonomic recognition with image processing techniques.

The Venice Lagoon (Northern Italy) is an attractive area of study due to its historical interest and ecological fragility. A spatial and temporal survey at three sites located in the "canals" of Venice city centre and at a reference site in the Lagoon was undertaken to evaluate stress effects on mussels sampled in Venice urban area, where raw sewage is discharged without treatment (57). A battery of biomarkers (metallothionein, micronuclei, condition index and survival in air) was used to evaluate the stress condition of the animals. At the same time, an alkali-labile phosphate assay (ALP) was performed in mussel hemolymph to find a biomarker of the estrogenic effect for this species. Biomarker results showed an impairment of the general health condition in the mussels coming from the urban area, in agreement with the chemical analysis.

Another study (58) surveyed the water quality in Venice urban areas in connection with the discharge of untreated sewage directly into canals, in addition to the pollutant load already present in these areas. One way of estimating the impact of these chemicals is the monitoring of the local fauna. In the search for good water quality indicators in Venice urban area, two physiological indices for mussels (Mytilus galloprovincialis) - survival in air and condition index - have been evaluated. Native mussels and also those transplanted into the urban area showed reduced survivability in air and decreased condition index values, indicating a less healthy status in animals collected from the urban canals. Data are discussed in relation with pollutant bioaccumulation.

Coastal environments are highly variable on a daily scale. In these environments, benthic foraminifera, a class of marine Protista, can be used as bioindicator (59). These organisms can define the extent of similarenvironmentalconditions(biotopes) through the study of the structure of their assemblage (presence – absence - relative dominance). The monitoring results show the capacity of the benthic foraminifera to monitor the changes occurring in unstable environments and to indicate the evolutionary trends of transitional environments.

The use of molecular techniques such as PCR assay (Polimerase Chain Reaction) for the determination of the genotoxic effects induced by the pollutants in the monitoring programs is rather recent.

This tool was employed in an assessment

study on recharged aquifers (60): this practice presents advantages for integrated water management in the anthropogenic cycle, namely, advanced treatment of reclaimed water and additional dilution of pollutants due to mixing with natural groundwater. Nevertheless, this practice represents a health and environmental hazard because of the presence of microorganisms pathogenic and chemical contaminants. In this study, the groundwater recharge systems in Torreele, Belgium, Sabadell, Spain, and Nardo, Italy, were investigated for fecal-contamination bacterial pathogens. indicators. and antibiotic resistance genes over the period of one year. Real-time quantitative PCR assays for Helicobacter pylori, Yersinia enterocolitica, and Mycobacterium avium subsp. paratuberculosis, human pathogens with long-time survival capacity in water, and for the resistance genes (ermB, mecA, blaSHV-5, ampC, tetO, and vanA) were adapted or developed for water samples by different impacts. characterized The resistance genes and pathogen concentrations were determined at five or six sampling points for each recharge system. The three aquifer recharge systems demonstrated different capacities for removal of fecal contaminants and antibiotic resistance genes.

A targeting species-specific PCR assay was combined with a filter system to collect phytoplankton cells and get spatial and temporal series as part of the Mediterranean Sea EU project Strategy (61). The application of PCR allowed a rapid detection of several harmful dinoflagellate species and genera, including *Alexandrium* spp. Field samples were concentrated on filter membranes, total DNA was extracted from mixed phytoplankton populations and PCR assays were carried out with specific primers. Qualitative PCR results were compared with light and epifluorescence microscopic examinations. Results indicated that this molecular assay was able to detect harmful target cells at concentrations undetectable by microscopy. The application of this filter PCR assay to seawater samples showed to be a sensitive and rapid procedure for the routine monitoring of coastal waters.

3.3 Ecological community studies and microbiology

The classification of the quality status of a water body through the study of the resident ecological community at different tiers is currently widely diffused in European monitoring programs also thanks to the impulse received by WFD. The effect of stressors on the ecological community is evaluated both at structure and function levels. Macrophytes, diatoms, macrobenthos and fish are the mostly used ecological quality indicators.

Among the other biological components of the ecological community in both water bodies and soils, special attention should be paid to the microbiological community for its functional role, ubiquitous presence and possible direct impact on human health.

In fact, the study on the natural capacity of bacterial communities to remove xenobiotics (pesticides, pharmaceuticals, biocides) in soil and water could have a very important implication for human health protection. The CNR-IRSA is currently studying natural bacterial communities from contaminated sites: specific bacteria strains that, after repeated exposure to xenobiotic, adapted themselves and became able to remove pollutants through metabolic and/or cometabolic processes, were isolated and identified (62-67). This "self-purification" ability can be used for "recovery strategies" (i.e. bioremediation) of contaminated sites.

The knowledge of the microbiological quality of coastal waters and marine organisms (fish, crustaceans and molluscs) (68-70), that is fundamental to assess the sanitary and ecological risk in a coastal zone, is a central research issue for the CNR Institute of Coastal Marine Environment (CNR-IAMC).

A recent work by Caruso et al. (71-75) is focused on the use of bacterial indicators to assess the anthropogenic pressures over coastal aquatic environments. Selected bacterial species (*Escherichia coli*, *Enterococcus* spp.) or related parameters allow to track the occurrence and evolution of bacterial pollution, and to prevent human health risks caused by the use of polluted waters.

Up-to-date standard procedures for bacterial pathogens determination and identification are necessary, due to the limitations of the usual culture methods (long response times, low specificity). In recent years, research efforts have been devoted to the improvement of technical equipment (automatic multiple samplers) and methodologies for the assessment seawater microbiological of quality particularly addressing the detection and enumeration of Escherichia coli or Enterococcus spp. in seawater as faecal pollution indicators. Rapid methods such as the fluorescent antibody method and the β-glucuronidase assay have been developed and optimized to monitor bathing waters, allowing the quantitative measurement of target bacterial molecules and accurate quantification faecal pollution of phenomena. Combined fluorescencestaining protocols have also been set up, in order to detect bacteria which may present a pathogenic potential. Data obtained by these new analytical procedures encourage the use of E. coli or related parameters

as successful tools for early warning of seawater bacterial pollution and for the screening of polluted coastal areas; therefore they offer interesting perspectives to prevent waterborne diseases.

The microbiological quality of coastal waters is usually estimated by determination of faecal indicator bacteria. bacterial However. species possibly pathogen for humans could occur also as microorganisms indigenous to coastal marine environments (e.g. Vibrio spp.). Since their concentration is related to the temperature of coastal waters, and unrelated to classical faecal indicators. monitoring and control of this bacterial group is needed to plan preventive measures for human health protection. (70,76-78) The Vibrio genus is widespread in coastal waters and includes more than 63 species. The most well-known species is V. cholerae, which causes cholera epidemics worldwide. In addition to V. cholerae, many other Vibrio spp. are recognized as potentially human pathogens causing 3 major syndromes of clinical illness: gastroenteritis, wound infections. and septicemia. Epidemiologic data suggest that the majority of these infections are foodborne disease and associated with raw or undercooked shellfish. In wild and cultivate shellfish the bacterial density may reach high concentration and potential toxic effect for humans. For this bacterial group the standard microbiological criteria used to assess water quality have to be revised.

The analysis of antibiotic resistance patterns (ARPs) of faecal indicator bacteria allows to detect the presence and persistence in the environment of genes linked to antibiotic resistance. Research is in progress to characterize the ARPs of enterococci, as they are emerging pathogenic bacteria, and of *E. coli* for their capacity of acquiring antibiotic resistance and spreading their resistance to other species *such as Salmonella, Shigella, Yersinia, Vibrio etc.* (79-81).

Studies on *E. coli and enterococci* have particular epidemiological and ecological relevance because these micro-organisms can occupy multiples niches, including humans, mammals, birds, reptiles and insects.

3.4 Advanced technologies and early warning systems

Advanced technologies have been applied to the monitoring of soil and water for the protection of human and environment health.

Among those, interest is growing for the application of innovative microscopy techniques. as scanning probe such microscopy (SPM). atomic force microscopy (AFM) and scanning near-field optical (SNOM) to the environmental field. These techniques exploit the interaction of a functional tip scanned over the surface of a sample (solid or adhesive on the substrate) to reconstruct the morphology of the sample and, simultaneously, achieve superresolution maps of other properties of the sample (i.e. local optical and fluorescence properties, maps of surface friction and nano-mechanical properties, magnetic domains, etc.). These microscopes are very useful in the study of nanostructured materials, surfaces and interfaces (i.e. biocompatible biosensors. sensors. surfaces), and analysis of nanoscale biosystems that can be investigated at the cellular level, subcellular and macromolecular aggregates. In particular, biosystems can be studied in physiological conditions and can describe temporal, three-dimensional and quantitative evolution (morphometry). (82-88) In addition to the development of innovative tools, expertise has been gained over the years in the study of intra-cells interactions in culture and exposed to environmental agents (drugs, electromagnetic fields, heavy metals, UV etc.) or nanostructured agents (nanoparticles or nanotubes) and in the study of complex phenomena such as cellular aging or apoptosis.

An important research field is the integration of advanced technologies in measuring systems that can be used for on-line or in-situ monitoring. The final goal is to make available some devices that can act as an early warning system of sudden alteration in the environmental quality status.

In the CNR-IRSA, Liquid Chromatography with Mass Spectrometric detection (LC-MS) has been integrated with an automatic sampling and preparation station into an on-line monitoring station for the determination of polar organic substances in drinking water (89). This system can be used in potable water treatment plants for the control of influent and effluent. In this way, a control on the incoming water quality and efficiency of the treatment procedure should be achieved. The operating limitations of this station are linked to the total cost of the equipment, that is still too high for a massive deployment, the reduced frequency of sampling and the need for highly qualified professionals for the frequent maintenance required.

Spectroscopic techniques are suitable to be integrated in small portable or fixed station devices for in-situ monitoring by optical fiber detection.

The concentration of several pollutants, usually present in industrial waste waters, could be predicted by the neural network data processing of absorption and fluorescence measurements in the visible spectral range. Proper network tuning provides quantitative analysis of many pollutants with sub-ppm resolution. Compact optical fiber instrumentation for absorption spectroscopy and an innovative flowcell for fluorescence measurements enable cost-effective, in situ, nonstop monitoring of waste waters (90).

In the framework of the development of new methods for measuring and monitoring soil pollution, the use of magnetic susceptibility as a proxy variable for monitoring heavy metals in soils has been explored (91). Magnetic measurements are carried out by using a magnetic susceptibility meter with two different probes for in situ field surveys. The relationships between heavy metal levels and magnetic susceptibility values of soil samples were assessed. Results suggest that a careful check of the experimental procedure plays a crucial role for using magnetic susceptibility measurements in situ monitoring of heavy metals.

Environmental management needs some tools capable of providing, over a relatively short time, integrated responses regarding the levels of contamination and the ecological consequences on different compartments of the concerned ecosystems. The biosensor "tool", that responds precisely to this necessity, consists of a biological active element - from an isolated enzyme to a whole organism - immobilized on a transducer system (sensor) for the selective and reversible determination of the presence and/or the concentration of certain chemical molecules in a sample. In fact selectivity and sensitivity, together with the possibility to have a portable tool are the main advantages of biosensors.

A compact and portable sensing device that combines the production and detection of hydrogen peroxide in a single flow assay has been proposed for herbicide detection in water (92). The principle on which the biosensor is based is that herbicides, under illumination, can inhibit the photosystem

electron transfer. Photosynthetic Π membranes isolated from higher plants photosynthetic micro-organisms, and immobilized and stabilized, can serve as a biorecognition element for a biosensor. The inhibition of photosystem II causes a reduced photoinduced production of hydrogen peroxide, which can be measured by a chemiluminescence reaction with luminol and the enzyme horseradish peroxidase.

Systems that use biological responses to detect environmental quality changes in continuous (on-line monitoring) in a simple, quick and economical way can be used as Biological Early Warning System. In this field, the possible use of an electrochemical growth signal transduction of a natural biofilm (microecosystem) on a suitable metallic substrate, was recently investigated with the aim of revealing, in real time, any alteration of the normal development of the microbial community induced by the presence of toxic substances in the aquatic environment (93-95). The prototype of the innovative patented biosensor shows that the biological electrochemical signal is significantly inhibited in the short term (minutes-hours) by known concentrations of a series of reference toxics.

Among the most advanced early warning systems, based on in vivo response of organisms to toxic agents, a Swimming Behavior Recorder System - able to measure the swimming behavior of marine invertebrates larvae exposed to toxic substances and/or environmental matrix under controlled conditions - has been presented (96-99). The methodology for the detection of alteration in swimming uses a prototype system to analyze video graphics, capable of automatically recording the aquatic organisms swimming speed, and providing, on the basis of the alteration of this parameter compared with a control, two toxicological endpoints: immobilization (acute) and alteration of swimming speed (sub-lethal).

3.5 Auxiliary techniques for monitoring

Geophysical techniques have been used for several years to measure hydraulic parameters in the monitoring and control of groundwater contamination. The most common techniques are the electromagnetic ones that are more sensitive to any changes in physical parameters (i.e. electrical conductivity) of soils and sub-soils caused by the presence of a particular contaminant in the water tablet or soil porosity. Therefore, the use of geophysics monitoring systems should be able to follow over time and space areas the evolution of a particular case of pollution, such as, for example, a leakage from a dumping site. The study and application of these non-invasive and lowcost technologies, integrated with the more traditional ones (sampling and diagnostic studies), will lead to a knowledge of the land and the environment, in order to provide a better safeguard level and to plan remediation procedures.

In the Hydrogeosite laboratory of the CNR Institute of Methodologies for Environmental Analysis (CNR-IMAA) a simulation plant has been built to perform hydrogeophysical experiments for the integrated study of contaminated soils and subsoils with the aim of creating a standard methodology for practical intervention (100-102).

The activity carried out by the CNR Research Institute for Geo-hydrological Protection (CNR-IRPI) in the field of erosion and hydro-meteorological monitoring is of fundamental importance for the protection of the water bodies and their basin.

4. Emerging issues

The literature review highlights that the monitoring focus is still mainly addressed to some classes of well studied molecules, such as trace metals and persistent organic pollutants, which represent a typology of pollution that emerged several decades ago, but for which there is still concern because of their persistence in soil and water sediments and their capacity to accumulate in the trophic chain.

However there are many relatively new classes of pollutants, which we daily deal with, but are not still regulated by legislation. For these classes, fundamental research and appropriate monitoring plans are needed in order to understand the environmental distribution and fate, which are necessary to assess the effective risks for ecosystem and humans. CNR Institutes are carrying out research on many of these emerging environmental issues and, among them, two classes can be chosen as case studies, i.e. the engineered nanoparticles and the perfluorinated compounds. In the following sections a short critical review of the knowledge gaps will be presented in order to list future research needs for these two emerging classes of compounds.

4.1 Engineered nanoparticles

Anthropogenic nanoparticles (engineered nanoparticles; ENPs) are used in nanotechnology to create products used in various fields such as agriculture, electronics, biomedicine, manufacturing, pharmaceuticals cosmetics industry. The use of nanomaterials containing ENPs is expected to continuously increase in the near future. The nanometer size range (0-100 nm) means that nanomaterials exhibit properties and functions other than those owned by the same materials with larger diameter and these properties are

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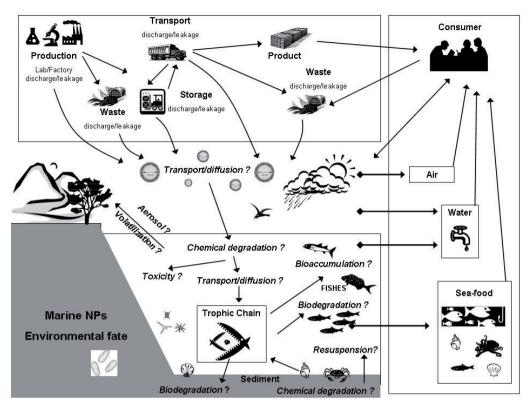


Figure 2: ENPs' environmental fate (figure adapted from National Institute for Resources and Environment, Japan http://www.nire.go.jp/eco_tec_e/hyouka_e.htm).

attributable to the increase of the surface area ratio and number per unit mass of ENPs, which lead to increased chemical reactivity, greater strength and electrical conductivity and, potentially, a more pronounced biological activity.

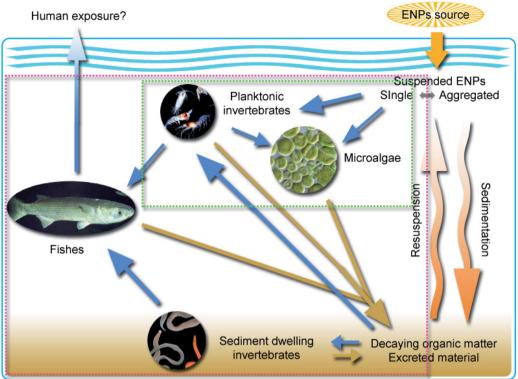
The extensive use of anthropogenic nanomaterials in large consumption products means that their transport, use and discharge is a potential new source of pollution in domestic sewage and industrial discharges, resulting in a diffused pollution of surface waters and transitional/coastal areas by ENPs. ENPs can be transferred to humans through food and several studies (103-120) demonstrate their potential risk for human health.

Today there are still many uncertainties about the environmental fate and toxicity of ENPs in aquatic environments because the methods for their determination and study in contaminated environments are not yet standardized.

The ENPs' environmental fate is extremely complex and the processes regulating ENPs' distribution in the different compartment are still not exhaustively investigated (Fig. 2). There are physico-chemical processes that can affect their potential environmental toxicity (solubility, aggregation, absorption, interaction with other toxic substances). The study of ENPs physico-chemical behavior influenced by abiotic factors in different matrices is fundamental to simulate scenarios in laboratory experiments.

According to Moore, some open issues can be suggested for future research projects(121):

- Which is the hydrodynamic behaviour of ENPs?
- Do they behave like larger natural particles?
- How do they associate with larger sediment



Mesocosm simulation of ENPs possible routes in the marine environment

Figure 3: Different levels of biological organization in a laboratory mesocosm (by Dr. Giovanni Pavanello, ISMAR).

and natural colloidal particulates?

- Do they bind lipophilic organic and metal pollutants, and what are the routes of nanoparticle uptake into biota?
- Do ENPs-linked chemical pollutants show enhanced toxicity?
- Are the particle size and surface properties significant factors in determining toxicity in aquatic organisms?
- What is the implication of ENPs exposure for organisms health and ecosystem integrity?
- Can modelled fluxes and predicted impacts of ENPs help to provide an explanatory framework for their environmental behaviour and possible impacts?

All these gaps could be filled by planning an appropriate laboratory experimental model which should involve different expertises available in the CNR Institutes, with the cooperation of other national and international research institutions.

For the evaluation model, different classes of anthropogenic ENPs representative of potential future scenarios of use and impact in the aquatic environment should be chosen. Their interaction with various compartments of aquatic ecosystems (water and sediment), in varying chemical and physical conditions (pH, temperature, ion concentration, natural matter concentration) should be determined.

For each ENPs' selected class a standardized protocol for measurement must be developed for each different environmental matricx (water, sediment and biota), including all the possible chemical and physical interactions between ENPs and the different matrices. One approach could be a careful and critical

analysis/testing of ENPs' "detection" methods already used in biomedical field (Environmental Scanning Electron Microscope-ESEM, **Field-Emission** Gun-Environmental Scanning Electron Microscope-FEG-ESEM, Scanning Transmission Electron Microscope-STEM, Transmission Electron Microscope-TEM, Microscope-SEM, Electron Scanning Atomic Force Microscope-AFM, Scanning Microscope-SPM, Probe Confocal Laser Scanning Microscope-CLSM and MultiPhoton Microscope-MPM) in order to adapt them to the specific peculiarities of the environmental matrices.

A crucial issue for a reliable risk assessment on ecosystems is the selection and standardization of ecotoxicological bioassays at different levels of biological organization (molecules, cells, organisms), which must be representative of different aquatic environments and ENPs categories selected as references standards. The selection of model species should be large enough to ensure a good representation of the different compartments of the most representative reference ecosystems.

Although it is reasonable and practical to predict the potential environmental risk on the basis of bioassays with single species models and/or battery of them, it is equally important to investigate these supramolecular entities how interact at ecosystem level. For this purpose, investigation protocols and exposure methods using realistic ENPs concentrations for various environments should be developed using laboratory micro and mesocosms containing sediment, microorganisms, algae, invertebrates and vertebrates, with different levels of ecosystem complexity. (Fig. 3).

The mesocosm study can also supply information on the accumulation and eventual biomagnification along the trophic chain. These data, together with human dietary and environmental exposure studies, are needed to get a reliable risk assessment for human health.

4.2 Perfluorinated compounds in the water environments

Perfluorinated alkyl substances (PFASs) are fluorinated man-made chemicals with unique molecular properties, chemical and thermal stability, water- and fat-repellent properties that make them and their derivatives useful in different industrial and household applications. They are widely employed for impregnation of textiles and paper, for cleaning aids, for fire fighting foams, for metal surface treatment and in the production of fluoropolymers.

Among PFASs, perfluorooctanoic (PFOA) and perfluorooctanesulfonic (PFOS) acids are two persistent and bioaccumulative end-stage metabolites of particular concern for human health and the environment due to their potential toxicological effects on animal and also human endocrine system; to date, there is no evidence of biodegradation of these chemicals.

Recent concerns with the toxic effects of PFOS and PFOA began in the early 2000s when the 3M Company, U.S.A, the major perfluorochemical manufacturer, decided to phase out the production of PFOS related products. However, there are still a number of industries, such as the semiconductor etc, that still use PFOS in their production (122) and manufacturers that use and produce PFOA and its related products for consumer usage. Evidence is accumulating regarding their persistence in the environment, potential for long-range transport, tendency to bioaccumulate, and potential toxicological effects (123-125). Due to their physico-chemical properties, these chemicals leaked into the water, they accumulate in surface waters, and water is the major reservoir of perfluorinated compounds (PFCs) in the environment, as well as the most important medium for their transport (126,127). Unlike other classical persistent organic pollutants (POPs), the anphiphilic PFCs have not been shown to accumulate preferentially in adipose tissue. They rather bind to blood proteins and accumulate in the liver of exposed organisms (128). Recent studies have indicated that PFOS can biomagnify in higher trophic organisms through the aquatic food chain (129,130).

PFOA is mainly used as adjuvant in the production of fluoropolymers such polifluorotetraethylene and similar as products used in clothing production, cosmetics and for non-stick cookware coatings. It was estimated that aqueous and gaseous emissions of PFOA or their salts originated in this production represent the majority (about 80%) of environmental release (126). It is also used (but in small amount) in industrial and consumer products (varnishes, inks, paper, floor polishes, cleaning formulation etc.) which represent some direct human exposure and diffusive environmental routes sources not normally considered. PFOA may also form as a degradation product of fluorotelomer alcohols found in a wide range of household and consumer products like hair shampoo, rug cleaners, and food paper products. These are volatile and can be carried for long distances by air currents (131).

In a European study on the major EU rivers carried out by the Perforce consortium (PERFORCE project), river Po watershed was identified as the dominant source of PFOA in Europe. River Po accounted for two thirds of the total PFOA discharge of the rivers studied (127). The PFOA concentration (200 ng/L) in the water of the Po river collected at the basin closure (Pontelagoscuro near Ferrara) was from 10 to 200 times higher than those measured in the other European rivers. This result suggested is the presence of an industrial source of PFOA in the Po watershed. Recent survey of perfluorocarboxylic and perfluorosulfonic acids concentrations in the Po watershed (33,132) confirmed the data of the PERFORCE project, and measured a PFOA concentration of 60-174 ng/l at Pontelagoscuro (FE). All tributaries, except river Tanaro, showed the PFOA levels typical of a diffuse pollution; on the contrary, the highest PFOA concentrations were determined in river Tanaro (1270 ng/L) and in river Po, downstream the confluence of the Tanaro river (60-337 ng/L), suggesting the presence of a point source for this compound that can be reasonably attributed to an industrial Teflon production site in the Alessandria county.

In the study carried out at the CNR-IRSA (33) an assessment of the contamination by perfluoroalkyl acids (PFAAs) in surface, urban and industrialized waste waters and drinking waters in the river Po basin was carried out. An HPLC method with ion trap mass spectrometry detection was developed for the analysis of perfluorinated carboxylates (from C₅ to C₁₀) and perfluorinated sulfonates (C₄ and C₈) with a LOQ of 2 ng/L. Water sample extraction was performed on weak anion exchange (WAX) cartridges.

In this work three kinds of waste waters, from textile industry discharge, urban wastewater treatment plant (WWTP) and mix urban-industrial WWTP, were analyzed because the industrial activity is considered the main source of PFOA and other homologues. PFOS was found at lower concentrations than PFOA, both in industrial and urban waste waters, This is probably due to the phase out of the

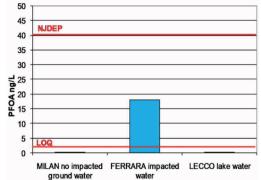


Figure 4. Perfluorinated concentration in tap waters. LOQ = Limit of quantification (2 ng/L); NJDEP = New Jersey Department of Environmental Protection guidance level (40 ng/L)

PFOS and PFOS-related compounds in the electrochemical production since 2002.

Three sources of tap waters (Lecco produced from Lake of Lecco water; Milano - produced from ground water; Ferrara - partially produced from the Po river) were also analyzed. The only perfluorinated compound, detected in drinking water above the limit of quantification, was PFOA in the tap water of Ferrara which produces drinking water mixing ground and surface waters from river Po (Fig. 4).

To date, there are no international water guidelines for PFOA in drinking water but some single countries set up their own water guidelines. The strictest standard, (40 ng/L), issued by New Jersey in the USA, was not overcome by Ferrara tap waters. Nevertheless it is necessary to underline that this sampling campaign on drinking waters does not have any statistical representativeness, but it is only a baseline survey which needs further in depth enquiry.

Preliminary results highlighted the presence, in the Poriver watershed, of point and diffused sources of perfluorinated compounds. PFAAs were measured in drinking waters produced from contaminated surface waters, revealing the risk for human consumption.

These preliminary campaigns, carried out in different periods and hydrological regimes, gave only a first look at the distribution of perfluorinated compounds in the Po watershed but no data have been collected for the rest of Northern Italy whose waters discharge in the High Adriatic Sea. The available data do not allow to estimate the risk of contamination for the aquatic species present in the transitional and coastal areas.

Furthermore the presence of intensive aquaculture activities in these areas could result in a source of exposure also for humans which, at the present state of knowledge, can not be envisaged.

Therefore, the overall goal of the IRSA research group will be to assess the environmental and health risks posed by the PFCs contamination in the river Po basin and in other river basins and coastal areas of the Northern Adriatic sea. In order to reach these objectives, the following detailed actions should be carried out:

- a comprehensive survey of the river Po basin , the coastal areas and lagoons of the High Adriatic Sea in order to identify critical areas and hot spots for PFC impact by:
 - Monitoring the river Po basin with downstream sampling of the main tributaries;
 - Evaluation of PFCs distribution in surface and groundwater around fluorochemical industrial sites in Northern Italy;
 - Monitoring of mollusc aquaculture areas in hot spot areas such as the Po Delta and Northern Adriatic Sea coast
 - Monitoring of raw and treated drinking water drawn from the Po river or groundwater impacted by industrial activities
- intensive monitoring of the critical areas in different hydrological regimes for the assessment of contamination and transport

processes;

- determination of PFCs levels in edible marine and lagoon bivalves, farmed or caught in the critical areas, in order to determine:
 - PFCs accumulation in different sites, using two edible species with different geographical distribution (site comparison);
 - PFCs accumulation in two different bivalves, addressing the possible existence of different patterns of exposure (clams in the sediment/mussels in the water column) and of bioaccumulation (species comparison);
- Estimation of PFCs risk exposure for humans as a result of the consumption of contaminated mollusks and drinking waters.

5. FUTURE PERSPECTIVES AND DEVELOPMENTS

The term "emerging pollutants" means a continuously updated list of the most diverse molecules produced, used and eventually diffused in the environment. Such list includes synthetic surfactants, disruptors, pharmaceutical endocrine substances, perfluorinated compounds, additives, new generation industrial pesticides, etc, all substances which are not included in the lists of regulated priority pollutants and in ordinary monitoring plans. Considering also that the toxicity and the effects of certain new generation compounds are often not known or fully understood, it is evident how crucial it is to develop monitoring systems suitable to detect the possible responses to unknown toxic agents. In fact it is impossible to measure all the compounds present in a certain environmental compartment and we need to turn over our traditional approach in monitoring: first we measure the effects and then we look for the molecular agents which should be the cause of the effect itself. This revolution in the monitoring approach will be probably the most outstanding innovation in the next future.

At present, the procedures identified by the words "Toxicity Identification and Evaluation" (TIE) or "Effect Directed Analysis" (EDA) represent the state of the art in this field and for that reason we will present a short review on this topic in this last section.

process The first identification of "guilty" substances and its verification of the observed effects caused by these compounds dates back to 1983 (133). Subsequently taken over by the United States Environmental Protection Agency (US EPA), this process took the name of Toxicity Identification Evaluation (TIE). The classic TIE is divided into 3 phases. During the first phase (Fig. 5a), groups of active compounds are sequentially removed from the water sample with chemical treatment, as long as the toxic effect of the solution on the biological model system does not disappear. In the second phase substances are tentatively identified by the treatment that decreases the toxic response of the sample and the classes of chemicals identified are related to the biological responses. Finally the activity of each identified substance or substances mixture is confirmed using the same biological model.

The Effect Directed Analysis (EDA) procedure, which is the subject of EU research projects in the framework of WFD implementation, is the conceptual and technical evolution of the TIE approach. Like TIE, EDA is based on the use of the response of a biological reference system in combination with sample fractionation and the chemical analysis of individual fractions for identification of hazardous compounds in complex mixtures present in the environment. Figure 5b schematically

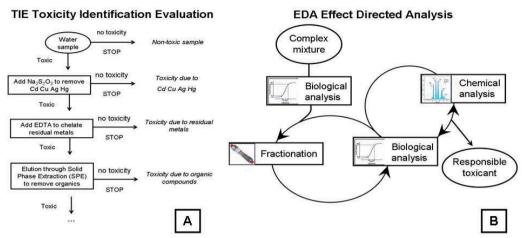


Figure 5: a-An example of TIE's first phase. b-scheme of EDA process.

describes the process which starts from the observation of a "macroscopic" effect on the biotic component in the investigation site. Through the use of a chromatographic separation systems, the mixture is divided into its individual components. Each fraction is tested for activity on in-vitro model systems, the nature of the compound is defined through analytical detection and recognition systems to determine the substance responsible for the effect.

fractionation into individual Sample classes of compounds is achieved using chromatographic columns with different stationary phases based on the principle that each compound has a specific affinity with the stationary phase column at a certain mobile phase composition. Therefore different types of serially connected chromatographic columns allow a sample basic components separation efficiency greater than that obtained using a single column, as in simpler TIE procedures (134,135).

The identification and confirmation of compounds in each fraction is achieved by using high-resolution mass spectrometry techniques, hyphenated with chromatographic systems; liquid chromatography (HPLC) is used for the separation and identification of polar compounds and metabolites which are more difficult to analyze but are much more biologically active than classic persistent organic pollutants.

In addition to a more rigorous procedure and a greater degree of complexity and precision in the sample fractionation compared to TIE. EDA introduces the use of in-vitro bioassays that allow a high number of spatial and temporal replicates and provide the capacity to run the tests in batteries, so getting a complete screening of the toxic activities of the sample fraction. As examples of in vitro tests used in EDA approach we can mention Yeast estrogen screen (YES), Yeast androgen screen (YAS), for the evaluation of endocrine disruption activity (136), bioluminescence tests such as CALUX, Microtox and Mutatox, assays on fish hepatocytes for physiological and accumulation studies: these are all sensitive, specific and rapid tests which allow to provide a full view on the biological activity of the compounds under investigation.

In parallel to those "classic" in-vitro tests, in recent years, the development of the "omics" disciplines helped to expand knowledge about the effects of many toxic compounds: toxicogenomics and toxicoproteomics study the relationships between genome (or protein) structure and activity and the biological effects of exogenous agents. This means that if a compound shows, for example, a mutagenic effect, the magnitude of the effect can be determined by the gene expression analysis of an exposed individual.

As an example of the effectiveness of the integrated monitoring and analysis approach, we can mention the study of Keiter et al. (137) who adopted an integrated TRIAD and EDA method to identify the activity of so-called dioxinlike compounds (i.e. molecules whose steric and electronic structure is similar to those structural characteristics of dioxins which exert effects on biological systems) as the cause of a specific adverse effect on river organisms.

The EDA procedure is rather complex and time consuming and, of course, it is not suitable to meet the demand of fast and reliable techniques for early warning Nevertheless, if integrated systems. into routine monitoring programs as a complementary technique for investigative monitoring, it will be absolutely necessary in those situations where the toxic or other biological effect cannot be attributed to a known agent or source. Alternatively it could be used as a screening method to assess the risk of a particular pollution source that is crucial for an effective management of the territory with the aim of human health and environmental safeguarding.

With the contribution of WG2 participants: A. Barra Caracciolo (IRSA), E. Belluso (IGG) G. Caruso (IAMC), L. Da Ros (ISMAR), M. Faimali (ISMAR), M. Girasole (ISM), R. La Ferla (IAMC), M. Mancuso (IAMC), G. Mascolo (IRSA), V. Matranga (IBIM), L.S. Monticelli (IAMC), E. Morelli (IBF), P. Gualtieri (IBF), M. Pennisi (IGG), M. Polemio (IRPI), E. Rizzo (IMAA), E. Sacchi (IGG), R. Zaccone (IAMC).

With the contribution of: A. Baldi (Fondazione Italiana Endometriosi); E. Fommei (Pisa University and CNR-IFC), G. Iervasi, A. Pierini (CNR-IFC); S. Maffei, C. Vassalle (Fondazione "G. Monasterio" CNR-Regione Toscana, Pisa); C. Roscioli, S.Valsecchi, L. Viganò, D. Vignati (CNR-IRSA, Brugherio).

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References

- 1. SETAC Society of Environmental toxicology and chemistry. Technical issue paper: Ecological risk assessment. Pensacola (FL): SETAC 1997: 4p.
- Chapman PM. The sediment quality Triad approach to determining pollutioninduced degradation. Sci Total Environ 1990;97–98:815–825.
- Long ER, Chapman PM. A Sediment Quality Triad: Measures of sediment contamination, toxicity and infaunal community composition in Puget Sound. Mar Poll Bull 1985;16:405–415.
- 4. Spurgeon DJ, Ricketts H, Svendsen C, Morgan AJ, Kille P. Hierarchical responses of soil invertebrates (earthworms) to toxic metal stress. Environ Sci Technol 2005;39:5327–5334.
- Dagnino A, Allen JI, Moore MN, Broeg K, Canesi L, Viarengo A. Development of an expert system for the integration of biomarker responses in mussels into an animal health index. Biomarkers 2007;12:155–172.
- Viarengo A, Gastaldi L, Dagnino A, Capri F, Torrielli S, Pons G. An expert system assessing pollutant-induced stress syndrome in the earthworm *Eisenia andrei*. Proceedings of the Society of Environmental Toxicology and Chemistry (SETAC) 17th Annual Meeting; 2007 May 20–24; Porto, Portugal. Pensacola

(FL): SETAC. 184 p.

- Spurgeon DJ, Hopkin SP, Jones DT. Effects of cadmium, copper, lead and zinc on growth, reproduction and survival of the earthworm *Eisenia fetida* (Savigny): Assessing the environmental impact of point-source metal contamination in terrestrial ecosystem. Environ Pollut 1994;84:123–130.
- Dickerson RL, Hooper MJ, Gard NW, Cobb GP, Kendall RJ. Toxicological foundations of ecological risk assessment: Biomarker development and interpretation based on laboratory and wildlife species. Environ Health Perspect 1994;102:65–69.
- 9. Ehlers LJ, Luthy RG. Contaminant bioavailability in soil and sediment. Environ Sci Technol 2003; 37:295A– 302A.
- Semple KT, Doick KJ, Jones KC, Burauel P, Craven A, Harms H. Defining bioavailability and bioaccessibility of contaminated soil and sediment is complicated. Environ Sci Technol 2004;38:228A–231A.
- Semenzin E, Temminghoff EJM, Marcomini A. Improving ecological risk assessment by including bioavailability into species sensitivity distributions: An example for plants exposed to nickel in soil. Environ Pollut 2007;148:642–647.
- Crumbling DM, Lynch K, Howe R, Groenjes C, Shockley J, Keith L et al. Managing uncertainty in environmental decisions. Environ Sci Technol 2001;35:404A–409A.
- 13. McCarthy JF, Shugart LR. Biomarkers of environmental contamination. Chelsea Mich.: Lewis Publisher;1990.
- Hagger JA, Jones MB, Leonard DRP, Owen R, Galloway TS. Biomarkers and integrated environmental risk assessment: are more questions than answers? Integr Environ Assess Manag 2006;2:312–329.
- 15. Dagnino A, Sforzini S, Dondero F, Fenoglio S, Bona E, Jensen J et al. A "Weight-of-Evidence" Approach for the Integration of Environmental "Triad" Data to Assess Ecological Risk and Biological Vulnerability. Integrated

Environmental Assessment and Management 2008;4(3):314-326.

- Bretzel F, Calderisi M. Metal contamination in urban soils of coastal Tuscany(Italy)EnvironmentalMonitoring And Assessment 2006;118(1-3): 319-335.
- 17. Scerbo R, Ristori T, Stefanini B, De Ranieri S, Barghigiani C. Mercury assessment and evaluation of its impact on fish in the Cecina river basin (Tuscany, Italy). Environmental Pollution 2005;135(1):179-186.
- Bianchini G, Pennisi M, Cioni R, Muti A, Cerbai N, Kloppmann W. Hydrochemistry of the high-boron groundwaters of the Cornia aquifer (Tuscany, Italy). Geothermics 2005;34:297-319.
- Gonfiantini R and Pennisi M. The behaviour of boron isotopes in natural waters and in water-rock interaction. J Geochem. Exploration 2006;88:114-117.
- Pennisi M, Gonfiantini R, Grassi S, Squarci P. The utilization of boron and strontium isotopes for the assessment of boron contamination of the Cecina River alluvial aquifer (central-western Tuscany, Italy). Applied Geochemistry 2006;21:643-655.
- Pennisi M, Bianchini G, Muti A, Klopmann W. Behaviour of boron and strontium isotopes in groundwateraquifer interactions in the Cornia Plain (Tuscany, Italy). Applied Geochemistry 2006;21:1169-1183.
- 22. Facchinelli A, Magnoni M, Perrone U, Sacchi E. Post-depositional processes in lake sediments traced by heavy metals and radionuclides: a case study from Lake Sirio (Ivrea, Northern Italy). Materials and Geoenvironment 2005;52:31-33.
- 23. Sacchi E, Brenna S, Fornelli I, Genot S, Sale VM, Azzolina L et al. Analisi del contenuto in rame ed altri metalli nei suoli agricoli lombardi.Quaderni della Ricerca 2007:61,Regione Lombardia,111.
- 24. Scarciglia F, Sacchi E, Angelone M, Apollaro C, Armiento G, Barca D et al. Caratteri geochimici, isotopici e mineralogici dei suoli di Muravera. In: Ottonello G. (Ed.) Geochemical Baselines

of Italy, Pisa, Pacini Editore 2007;87-147.

- 25. Sacchi E, Brenna S, Fornelli Genot S, Setti M, Sale VM et al. A regional survey on heavy metals content in cultivated soils from Lombardy (Italy): results from the RAMET project. Int. Symp. "Consoil 2008", Milan (Italy), 2008 3-6 June 2008, C164-C172.
- Cardellicchio N, Buccolieri A, Di Leo A, Giandomenico S, Spada L. Levels of metals in reared mussels from Taranto Gulf (Ionian Sea, Southern Italy). Food Chemistry 2008;107,(2):890-896.
- 27. Belluso E, Fornero E, Cairo S, Albertazzi G, Rinaudo C. The application of micro-Raman spectroscopy to distinguish carlosturanite from serpentine-group minerals. Canadian Mineralogist 2007;45:1495-1500.
- Cardile V, Lombardo L, Belluso E, Panico A, Capella S, Balazy M. Toxicity and carcinogenicity mechanisms of fibrous antigorite. International Journal of Environmental Research and Public Health 2007;4:1-9.
- 29. Cardile V, Lombardo L, Belluso E, Panico AM, Renis M, Gianfagna A et al. Fluoro-edenite Fibers Induce Expression of Hsp70 and Inflammatory Response. International Journal of Environmental Research and Public Health 2007;20:195-202.
- Fornero E, Belluso E, Capella S, Bellis D. Environmental exposure to asbestos and other inorganic fibres using animal lung model. Science of the Total Environment 2009; 407(3):1010-1018.
- Detomaso A, Mascolo G, Lopez A. Characterization of carbofuran photodegradation by-products by liquid chromatography/hybrid quadrupole time-of-flight/mass spectrometry. Rapid Communication in Mass Spectrometry 2005;19:2193–2202.
- 32. Cavalli S, Polesello S, Saccani G. Determination of Acrylamide in Drinking Water by Large-Volume Direct Injection and ICE-MS detection. Journal of Chromatography A 2004;1039:155-159.
- 33. Valsecchi S, Polesello S. The search of

sources of perfluoroalkyl acids (PFAAs) in Northern Italian waters, 19th Annual Meeting of Setac Europe "Protecting ecosystem health: facing the challenge of a globally changing environment". Gőteborg, Sweden. 2009 May 31- June4.

- 34. Marin MG, Boscolo R, Cella A, Degetto S, Da Ros L. Field validation of autometallographical black silver deposit (BSD) extent in three bivalve species from the lagoon of Venice, Italy (*Mytilus galloprovincialis, Tapes philippinarum, Scapharca inequivalvis*) for metal bioavailability assessment. Sci Tot Environ 2006;371:156-167.
- 35. Da Ros L, Moschino V, Guerzoni S, Halldòrsson HP. Lysosomal responses and metallothionein induction in the blue mussel *Mytilus edulis* from the south-west coast of Iceland. Environ International 2007;33(3):362-369.
- 36. Neston N, Romano S, Moschino V, Mauri M, Da Ros L. Chemical analysis and biomarkers in mussels and fish as tools for evaluating presence and effects of microorganic pollutants and trace metals in the lagoon of Venice, Italy. Mar Poll Bull 2007;55: 469-484.
- Angelini C, Amaroli A, Falugi C, Di Bella G, Matranga V. Acetylcholinesterase activity is affected by stress conditions in Paracentrotus lividus coelomocytes. Marine Biology 2003;143(4):623-628.
- Russo R, Bonaventura R, Zito F, Schroder HC, Muller I, Muller WEG et al. Stress to cadmium monitored by metallothionein gene induction in Paracentrotus lividus embryos. Cell Stress & Chaperones 2003;8(3):232-241.
- 39. Bonaventura R, Poma V, Costa C, Matranga V. UVB radiation prevents skeleton growth and stimulates the expression of stress markers in sea urchin embryos. Biochem Biophys Res Commun 2005,328:150-157.
- 40. Bonaventura R, Poma V, Russo R, Zito F, Matranga V. Effects of UV-B radiation on the development and hsp 70 expression in sea urchin cleavage embryos. Marine Biology 2006;149:79-86.

- Matranga V, Pinsino A, Celi M, Natoli A, Bonaventura R, Schröder HC et al. Monitoring chemical and physical stress using sea urchin immune cells. Prog Mol Subcell Biol 2005;39:85-110.
- 42. Matranga V, Zito F, Costa C, Bonaventura R, Giarrusso S, Celi F. Embryonic development and skeletogenic gene expression affected by X-rays in the Mediterranean sea urchin *Paracentrotus lividus*. Ecotoxicology [Epub ahead of print]Online: DOI 10.1007/s10646-009-0444-9. 2009 Nov 27.
- 43. Pinsino A, Matranga V, Trinchella F, Roccheri MC. Sea urchin embryos as an in vivo model for the assessment of manganese toxicity: developmental and stress response effects. Ecotoxicology [Epub ahead of print]Online: DOI 10.1007/ s10646-009-0432-0. 2009 Nov 1.
- 44. Schröder HC, Di Bella G, Janipour N, Bonaventura R, Russo R, Müller WE et al. DNA damage and developmental defects after exposure to UV and heavy metals in sea urchin cells and embryos compared to other invertebrates. Prog Mol Subcell Biol. 2005;39:111-37.
- 45. Filosto S, Roccheri MC, Bonaventura R, Matranga V. Environmentally relevant cadmium concentrations affect development and induce apoptosis of *Paracentrotus lividus* larvae cultured in vitro. Cell Biol Toxicol 2008;24: 603-610.
- 46. Pinsino A, Della Torre C, Sammarini V, Bonaventura R, Amato E, Matranga V. Sea urchin coelomocytes as a novel cellular biosensor of environmental stress: a field study in the Tremiti Island Marine Protected Area, Southern Adriatic Sea, Italy. Cell Biol Toxicol 2008; 24: 541-552
- 47. Morelli E, Scarano G. Copper-induced changes of non-protein thiols and antioxidant enzymes in the marine microalga *Phaeodactylum tricornutum*. Plant Sci 2004;167:289–296.
- 48. Loreti V, Toncelli D, Morelli E, Scarano G, Bettmer J. Biosynthesis of Cd-bound Phytochelatins (PCs) by *Phaeodactylum tricornutum* and their Speciation via

Size-Exclusion Chromatography (SEC) and Ion-Pairing Chromatography (IPC) coupled to ICP-MS. Anal Bioanal Chem 2005;383:398-403.

- 49. Morelli E, Mascherpa MC, Scarano G. Biosynthesis of phytochelatins and arsenic accumulation in the marine microalga *Phaeodactylum tricornutum* in response to arsenate exposure. BioMetals 2005;18:587-593.
- 50. Bramanti E, Toncelli D, Morelli E, Lampugnani L, Zamboni R, Miller KE et al. Determination and characterization of phytochelatins by liquid chromatography coupled with on line chemical vapour generation and atomic fluorescence spectrometric detection. J Chromatogr 2006;1133:195-203.
- 51. Morelli E, Fantozzi L. Phytochelatins in the Diatom *Phaeodactylum tricornutum* Bohlin: An evaluation of their use as biomarkers of metal exposure in marine waters. Bull Environ Contam Toxicol 2008;81:236-241.
- 52. Morelli E, Marangi ML, Fantozzi L. A phytochelatin-based bioassay in marine diatoms useful for the assessment of bioavailability of heavy metals released by polluted sediments. Environ Int 2009 (in press).
- 53. Evangelista V, Barsanti L, Passatelli V, Frassanito A, Gualtieri P. Microspectroscopy of the Photosynthetic Compartment of Algae. Photochem Photobiol 2006;82:1039-1046.
- Evangelista V, Evangelisti M, Barsanti L, Frassanito AM, Passarelli V, Gualtieri P. A polychromator - based microspectrophotometer". Int J Biol Sci 2007;3:251-256.
- 55. Rodriguez MC, Barsanti L, Passarelli V, Evangelista V, Conforti V, Gualtieri P. Effects of chromium on photosynthetic and photoreceptive apparatus of the alga Chlamydomonas reinhardtii. Environmental Research 2007;105:234-9.
- Barsanti L, Coltelli P, Evangelista V, Passarelli V, Rassanito AM, Gualtieri P. Low resolution characterization of the 3D structure of the *Euglena*

gracilis photoreceptor. Biochemical and Biophysical Research Communications 2008;375:471-476.

- 57. Pampanin DM, Marangon I, Volpato E, Campesan G, Nasci C. Stress biomarkers and alkali-labile phosphate level in mussels (Mytilus galloprovincialis) collected in the urban area of Venice (Venice Lagoon, Italy). Environmental Pollution 2005;136(1):103-107.
- 58. Pampanin DM, Volpato E, Marangon I, Nasci C. Physiological measurements from native and transplanted mussel (*Mytilus galloprovincialis*) in the canals of Venice. Survival in air and condition index. Comparative Biochemistry and Physiology A-Molecular & Integrative Physiology 2005;140(1):41-52.
- Albani A, Serandrei-Barbero R, Donnici S. Foraminifera as ecological indicators in the Lagoon of Venice, Italy. Ecological Indicators 2007;7(2):239-253.
- 60. Bockelmann U, Dorries HH, Ayuso-Gabella MN, de Marcay MS, Tandoi V, Levantesi C et al. Quantitative PCR Monitoring of Antibiotic Resistance Genes and Bacterial Pathogens in Three European Artificial Groundwater Recharge Systems. Applied and Environmental Microbiology 2009;75(1):154-163.
- 61. Penna A, Fusco G, Bertozzini E, Giacobbe MG, Vila M, Galluzzi L et al. Monitoring of Alexandrium species in the Mediterranean Sea using a combined filter system-PCR assay detection method. African Journal of Marine Science 2006;28(2):241-243.
- 62. Accinelli C, Barra Caracciolo A, Grenni P, Giuliano G, Vicari A. Degradation of the antiviral drug oseltamivir carboxylate in surface water. In: *Environmental Fate and Ecological effects of pesticides*. Del Re AAM, Capri E, Fragoulis G, Trevisan M Eds, La Goliardica Pavese 2007:401-407.
- 63. Grenni P, Barra Caracciolo A, Saccà ML, Falconi F, Ciccoli R, Ubaldi C et al. Natural attenuation capability of an agricultural soil to degrade terbuthylazine.

In: *Environmental Fate and Ecological effects of pesticides*. Del Re AAM, Capri E, Fragoulis G, Trevisan M Eds, La Goliardica Pavese 2007:601-606.

- 64. Martín M, Gibello A, Martínez-Iñigo MJ, Lobo MC, Nande M, Garbi C et al. Proposal for a Natural Attenuation Coefficient for simazine- contaminated soils based on fluorescence in situ hybridization. Chemosphere 2008;71:703-710.
- 65. Singer A, Howar M, Johnson A, Accinelli C, Bird S, Boucard T et al. Meeting report: risk assessment of Tamiflu® use under pandemic conditions - Report from an interdisciplinary Workshop. Environmental Health Perspectives 2008;116(11):1563 -1567.
- 66. Grenni P, Barra Caracciolo A, Rodríguez-Cruz MS, Sánchez-Martín MJ. Changes in the microbial activity in a soil amended with oak and pine residues and treated with linuron herbicide. Applied Soil Ecology 2009;41:2-7.
- 67. Grenni P, Gibello A, Barra Caracciolo A, Fajardo C, Nande M, Sacca ML et al. A new fluorescent oligonucleotide probe for in situ detection of s-triazinedegrading *Rhodococcus wratislaviensis* in contaminated groundwater and soil samples. Water Research 2009;accepted
- 68. Mancuso M, Avendaño-Herrera R, Magariños B, Zaccone R, Toranzo AE. Evaluation of different DNA-based fingerprinting methods for typing Photobacteriumdamselaesubsp.piscicida. Biol Res 2007;40(1):85-92.
- 69. Zaccone R, Mancuso M. First report on antibodies response of Seriola dumerilii (Risso 1810) challenged with *Listonella anguillarum*. Fish and Shellfish Immunology 2008;25(5):689-692.
- 70. Leonardi M, Azzaro F, Azzaro M, Caruso G, Mancuso M, Monticelli LS et al. Multidisciplinary study of the Cape Peloro brackish area (Messina, Italy): characterisation of trophic conditions, microbial abundance and activities. In: Marine Ecology: an evolutionary perspective (S.Z.N.). M.C. Gambi and

Levin eds; 2009. p.33-42.

- 71. Caruso G, Zappalà G, Crisafi E. Monitoring bacterial pollution in coastal waters: recent advances in technologies and rapid methods". 4th International Conference on Marine Waste Water Discharges and Marine Environment, Antalya (Turkey), 2006 November 6-10, Abstracts, 333-334.
- 72. Zappalà G, Caruso G, Azzaro F, Crisafi E. Marine environment monitoring in coastal Sicilian waters. In: Brebbia CA editor. Water Pollution VIII: Modelling, Monitoring and Management, Bologna, 2006 September 4-6;95:337-346. WIT Press, Southampton (UK).
- Caruso G, Crisafi E, Caruso R, Zappalà G. Advances in marine bacterial pollution monitoring. In: *Environmental Microbiology Research Trends*, G. V. Kurladze editor, Chapter 10, pp. 273-287, NOVA Publishers: Hauppauge, NY. USA; 2008.
- 74. Caruso G, Monticelli LS, Caruso R, Bergamasco A. Development of a fluorescent antibody method for the detection and enumeration of *Enterococcus faecium* and its potential for coastal aquatic environment monitoring. Marine Pollution Bulletin 2008;56:318-324.
- 75. Caruso G, Zappalà G, Maimone G, Azzaro F, Raffa F, Caruso R. Assessment of the abundance of actively respiring cells and dead cells within the total bacterioplankton of the Strait of Messina waters. In: Brebbia CA editor, Environmental Problems in Coastal Regions VII, The New Forest (UK), 2008 May 19-21. Proceedings, pp.15-24. WIT Press, Southampton (UK).
- Zaccone R, Azzaro M, Azzaro F, Caruso G, Giacobbe MG, Mancuso M et al. First microbiological data from the lagoon area of Cape Peloro (Messina). workshop 10 -131 Geoitalia 2007 Settembre 10-12.
- 77. Zaccone R, Azzaro M, Azzaro F, Caruso G, Mancuso M, Monticelli LS et al. Multidisciplinary study of Cape Peloro brackish area (Messina, Italy) 43 EMBS

Azzorre 2008 September 8-12; P3.42.

- Maimone G, Caruso G, La Ferla R, Mancuso M, Zaccone R. Variabilità della popolazione batterica in un ecosistema salmastro della Sicilia. II Workshop annuale VECTOR, 2009 February 25– 26; Roma.
- 79. Bergamasco A, Azzaro F, Caruso G, Crisafi E, Decembrini F, Monticelli LS et al. Tecniche e metodologie per la valutazione dello stato ecologico delle acque costiere e di transizione: risultati, strategie e prospettive. 1° Forum Istituto per l'Ambiente Marino Costiero, Giardini Naxos (ME), 6-9 maggio 2007.
- Caruso G, Zaccone R, Monticelli LS, La Cono V, Crisafi E. Impatto antropico su aree marine costiere: metodologie innovative per il controllo igienicosanitario delle acque. 1º Forum Istituto per l'Ambiente Marino Costiero, Giardini Naxos (ME), 6-9 maggio 2007.
- Caruso G, Monticelli LS, Caruso R, Bergamasco A. Development of a fluorescent antibody method for the detection and enumeration of *Enterococcus faecium* and its potential for coastal aquatic environment monitoring. Marine Pollution Bulletin 2008;56:318-324.
- Longo G, Girasole M, Cricenti A. A novel tapping SNOM: Instrument description and performances. Phys Stat Sol B 2005;242:3070-3074.
- 83. Moretti PF, Maras A, Palomba E, Girasole M, Pompeo G, Longo G. Detection of nanostructured metal in meteorites: implications for the reddening of asteroids. Astrophys J Lett 2005;634,L117-120.
- 84. Cattaruzza F, Cricenti A, Flamini A, Girasole M, Longo G, Prosperi et al. Controlled loading T. of oligodeoxyribonucleotide monolayers onto unoxydised crystalline silicon: fluorescence-based determination of the surface coverage and of the hybridisation efficiency; parallel imaging of the process by AFM. Nucleic Acid Research 2006;34(4):e32.
- 85. Colonna S, Pompeo G, Girasole M,

Gazzoli D, Pettiti I, Valigi M. Thermallyinduced morphological transition in WOx deposited on a ZrO2(100) substrate. Surf Sci 2007;601:1389–1393.

- 86. Girasole M, Cricenti A, Generosi R, Longo G, Pompeo G, Cotesta S et al. Different Membrane Modifications Revealed by Atomic Force/Lateral Force Microscopy After Doping of Human Pancreatic Cells With Cd, Zn or Pb. Micr. Res. Tech. 2007;70:912-917
- Girasole M, Pompeo G, Cricenti A, Congiu-Castellano A, Andreola F, Serafino A, et al. Roughness of the Plasma Membrane as an Independent Morphological Parameter to Study RBCs: a Quantitative Atomic Force Microscopy Investigation. Biochim Biophys Acta-Biomembranes 2007; 1768,1268–1276.
- Longo G, Girasole M, Cricenti A. Implementation of a bimorph-based aperture tapping-SNOM with an incubator to study the evolution of cultured living cells. J Microscopy 2008;229:433-439.
- Brumaru C, Polesello S, Valsecchi S. LC-MS-Ion Trap determination of natural and synthetic estrogens in drinking water, 1st thematic workshop of the EU project NORMAN – Chemical Analysis of Emerging Pollutants; 2006 November 27-28; Maò, Menorca (Balearic island), Spain; 2006. p. 101.
- 90. Kuzniz T, Halot D, Mignani AG, Ciaccheri L, Kalli K, Tur M et al. Instrumentation for the monitoring of toxic pollutants in water resources by means of neural network analysis of absorption and fluorescence spectra. Sensor and Actuators B-Chemical 2007;121(1):231-237.
- 91. D'Emilio M, Chianese D, Coppola R, Macchiato M, Ragosta M. Magnetic susceptibility measurements as proxy method to monitor soil pollution: development of experimental protocols for field surveys. Environmental Monitoring and Assessment 2007;125(1-3):137-146.
- 92. Varsamis DG, Touloupakis E, Morlacchi P, Ghanotaskis FD, Giardi MT, Cullen DC. Development of a photosystem II-based optical microfluidic sensor for herbicide

detection. Talanta 2008;77(1):42-47.

- 93. Garaventa F, Corrà C, Di Fino A, Modugno S, Mollica A, Pavanello G, et al. Automated systems to monitor marine pollution, from eco-toxicological kit to online Biological Early Warning Systems: an integrated approach. III Bilateral Seminar Italy-Japan on: Physical and Chemical Impacts on Marine Organisms - Seeking sustainability and postgenomics; Nagoya (Japan); 24-27 November 2008.
- 94. Faimali M, Chelossi E, Garaventa F, Corrà C, Greco G, Mollica A. Evolution of oxygen reduction current and biofilm on stainless steels cathodically polarised in natural aerated seawater. Electrochimica Acta 2008;54(1)148-15.
- 95. Pavanello G, Pittore M, Mollica A, Mollica A, Cappello M, Capparelli E et al. Sviluppo di un biosensore elettrochimico per la tossicità delle acque (beta), Biologia Marina Mediterranea 2009;in press.
- 96. Faimali M, Garaventa F, Piazza V, Greco G, Corrà C, D'Amico G. Mortality, settlement inhibition and swimming speed alteration of larvae of *Balanus amphitrite* as acute, chronic and behavioural end-point for laboratory toxicological bioassays. Biologia Marina Mediterranea 2007;14(1):114-116.
- 97. Garaventa F, Corrà C, Di Fino AL, Modugno S, Mollica A, Mollica A et al. Automated systems to monitor marine pollution, from eco-toxicological kit to online Biological Early Warning Systems: an integrated approach. III Bilateral Seminar Italy-Japan on: Physical and Chemical Impacts on Marine Organisms - Seeking sustainability and postgenomics – Nagoya (Japan), 24-27 November 2008.
- 98. Meneghetti F, Garaventa F, Bon D, Di Fino A, Gambardella C, Faimali M et al. Use of marine invertebrates behavioural endpoints to evaluate the toxicity of coastal sediment. International Expert Meeting on The Impacts of Human Activities at Sea, on The Coast and in Its Hinterland on The Northern Adriatic's Biodiversity – Piran (SI), October 7th – 8th, 2008.
- 99. Gambardella C, Di Fino A, Garaventa

F, Pittore M, Faimali M. Alterazione del nuoto larvale di organismi marini come end-point sub-letale in biosaggi ecotossicologici. Biol Mar Medit 2009; in press.

- 100. Suski B, Rizzo E, Revil A. A sandbox experiment of self-potential signals associated with a pumping test. Vadose Zone Journal 2005;3(4):1193-1199.
- 101. Bavusi M, Lapenna V, Rizzo E. Electromagnetic methods to characterize the Savoia di Lucania waste dump (Southern Italy). Environmental Geology 2006;51(2):301-308.
- 102. Straface S, Fallico C, Troisi S, Rizzo E, Revil A. Estimating of the transmissivities of a real aquifer using a Self Potential signals associated with a pumping test. Ground Water 2007;45(4):420-428.
- 103. Oberdòrster G, Gelein RM, Ferin J, Weiss B. Association of particulate air pollution and acute mortality: involvement of ultrafine particles? InhalToxicol. 1995;7:111-24.
- 104.
- 105. Marconi A. Particelle fini, ultrafini e nanoparticelle in ambiente di vita e di lavoro: possibili effetti sanitari e misura dell'esposizione inalatoria. G Ital Med Lav Erg 2006;28: 258-65.
- 106. Long TC, Saleh N, Tilton RD, Lowry GV, Veronesi B. Titanium dioxide (P25) produces reactive oxygen species in immortalized brain microglia (BV2): implications for nanoparticle neurotoxicity. Environmental Science and Technology 2006;40:4346-52.
- 107. Holsapple MR, Farland WH, Landry TD, Monteiro-Riviere NA, Carter JM, Walker NJ et al. Research strategies for safety evaluation on nanomaterials, part II: toxicological and safety evaluation on nanomaterials, current challenges and data needs. Toxicological Sciences. 2005;88: 12-7.
- 108. Magrini A, Bergamaschi A, Bergamaschi E. Nanotubi di carbonio (Ntc) e nanoparticelle (Np): interazione con i sistemi biologici con particolare riferimento all'apparato respiratorio. G

Ital Med Lav Eng. 2006;28:266-9.

- 109. Oberdòrster G, Maynard A, Donaldson K, Castranova V, Fitzpathck J, Ausman K et al. Principles for characterizing the potential human health effects from exposure to nanomaterials: elements of a screening strategy. Part Fibre Toxicol. 2005;2:8.
- 110. Oberdòrster G, Sharp Z, Atudorei V, Elder A, Gelein R, Kreyling W et al. Translocation of inhaled ultrafine particles to the brain. Inhal Toxicol. 2004;16:437-45.
- 111. Lockman RR, Mumper RJ, Khan MA, Alien DD. Nanoparticle technology for drug delivery across the blood-brain barrier. Drug Dev Ind Pharm. 2002;28:1-12.
- 112. Shvedova AA, Castranova V, Kisin ER, Schwegler-Berry D, Murray AR, Gandelsman VZ et al. Exposure to carbon nanotube material: assessment of nanotube cytotoxicity using human keratinocyte cells. J Toxicol Environ Health A 2003;66:1909-26.
- 113. Rahman Q, Lohani M, Dopp E, Pemsel H, Jonas L, Weiss DG et al. Evidence that ultrafine titanium dioxide induces micronuclei and apoptosis in Syrian hamster embryo fibroblasts. Environ Health Perspect. 2002;110:797-800.
- 114. Hansen T, Clermont G, Alves A, Eloy R, Brochhausen C, Boutrand JP et al. Biological tollerance of different materials in bulk and nanoparticulate form in a rat model: Sarcoma development by nanoparticles. J.R. Cos. Interface 2006;3:767-775.
- 115. Ballestri M, Baraldi A, Gatti AM, Furci L, Bagni A, Loria P et al. Liver and kidney foreign bodies granulomatosis in a patient with malocclusion, bruxism and worn dental prosthesis. Gastroenterology 2001;121.
- 116. Hansen T, Clermont G, Alves A, Eloy R, Brochhausen C, Boutrand JP et al. Biological tollerance of different materials in bulk and nanoparticulate form in a rat model: Sarcoma development by nanoparticles. J.R. Cos. Interface 2006;

3:767-775.

- 117. Gatti AM, Kirkpatrick J, Gambarelli A, Capitani F, Hansen T, Heloy R et al. ESEM evaluation of muscle/nanoparticles interface in a rat model. Mater Sci Mater Med 2008;19(4):1515-22.
- 118. Gatti AM, Balestri M, Bagni A. Granulomatosis associated to procelain wear debris. American Journal of Dentistry 2002;15(6):369-372.
- 119. Gatti AM, Tossini D, Gambarelli A, Montanari S, Capitani F. Investigation of the Presence of Inorganic Micro- and Nanosized Contaminants in Bread and Biscuits by Environmental Scanning Electron Microscopy. Crit Rev Food Sci Nutr. 2009;49(3):275-82.
- 120. Guildford AL, Poletti T, Osbourne LH, Di Cerbo A, Gatti AM, Santin M. Nanoparticles of a different source induce different patterns of activation in key biochemical and cellular components of the host response. J R Soc Interface 2009;6(41):1213-1221.
- 121. Bregoli L, Chiarini F, Gambarelli A, Sighinolfi G, Gatti AM, Santi P et al. Toxicity of antimony trioxide nanoparticles on human hematopoietic progenitor cells and comparison to cell lines. Toxicology 2009;262(2):121-9.
- 122. Moore MN. Do nanoparticles present ecotoxicological risks for the health of the aquatic environment? Environment International 2006;32:967–976.
- 123. Tang CY, Shang Fu Q, Robertson AP, Croddle CS, Leckie JO. Use of reverse osmosis membranes to remove perfluorooctane sulfonate (PFOS) from semiconductor wastewater. Environ Sci Technol 2006;40:7343-7349.
- 124. Hekster F, Laane RWPM, de Voogt P. Environmental and toxicity effects of perfluoroalkylated substances. Rev. Environ. Contam. Toxicol. 2003;179:99– 121.
- 125. Andersen ME, Butenhoff JL, Chang A-C, Farrar DG, Kennedy GL, Lau C et al. Perfluoroalkyl acids and related chemistries – Toxicokinetics and modes of action. Toxicological Sciences 2008;

102:3-14.

- 126. Condor JM, Hoke RA, de Wolf W, Russell MH, Buck RC. Are PFCAs bioaccumulative? A critical review and comparison with regulatory criteria and persistent lipophilic compounds. Environ Sci Technol 2008;42:995–1003.
- 127. Prevedouros K, Cousins IT, Buck RC, Korzeniowski SH. Sources, Fate and Transport of Perfluorocarboxylates. Environ Sci Technol 2006;40:32-44.
- 128. McLachlan MS, Holmstrom KE, Reth M, Berger U. Riverine Discharge of Perfluorinated Carboxylates from the European Continent. Environ Sci Technol 2007;41:7260-7265.
- 129. Jone PD, Hu W, de Coen W, Newsted JL, Giesy JP. Binding of perfluorinated fatty acids to serum proteins. Environ. Toxicol. Chem. 2003;22:2639-2649.
- 130. Martin JW, Whittle DM, Muir DCG, Mabury SA. Perfluorinated contaminats in a food web from Lake Ontario. Environ Sci Technol 2004;38:5379-5385.
- 131. Kannan K, Tao L, Sinclair E, Pastva SD, Jude DJ, Giesy JP. Perfluorinated compounds in aquatic organisms at various trophic levels in a Great Lake food chain. Arch Environ Contam Toxicol 2005; 48:559-566.
- 132. Ellis DA, Martin JW, De Silva AO, Mabury SA, Hurley MD, Sulbaek A et al. Degradation of fluorotelomer alcohols: a likely atmospheric source of perfluorinated carboxylic acids. Environ Sci Technol 2004;38:3316-3321.
- 133. Loos R, Locoro G, Huber T, Wollgast J, Christoph E, de Jager AM et al. Analysis of perfluorooctanoate (PFOA) and other perfluorinated compounds (PFCs) in the River Po watershed in N-Italy. Chemosphere 2008;71:306–313.
- 134. Samailoff MR, Bell J, Birkholz DA, Webster GRB, Arnott EG, Pulak R et al. Environmental Science and Technology 1983;17:329-334.
- 135. Brack W, Klamer HJC, Lopez M, Barcelo D. Effect-directed analysis of key toxicants in European river basin a review. Env Sci Pollut Res 2007;14(1):30-38.

CNR Environment and Health Inter-departmental Project

- 136. Brack W, Schmitt-Jansen M, Machala M, Brix R, Barcelo D, Schymenski E et al. How to confirm identified toxicants in effect-directed analysis. Anal Bioanal Chem 2008;390:1959-1973.
- 137. Viganò L, Mandich A, Benfenati E, Bertolotti R, Bottero S, Porazzi E et al. Investigating the estrogenic risk along the river Po and its intermediate section. Arch. Environ. Contam. Toxicol. 2006;51:641-651.
- 138. Keiter S, Grund S, Van Bavel B, Hagberg J, Engwall M, Kammann U et al. Activities and identification of aryl Hydrocarbon receptor agonists in sediment from the Danuberiver. Analytical and Bioanalytical Chemistry 2008;390:2009-2019.